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RESEARCH ARTICLE

4 **Impacts of urbanization on the distribution of heavy metals**
5 **in soils along the Huangpu River, the drinking water**
Q1 6 **source for Shanghai**

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9 Received: 24 May 2015 / Accepted: 3 November 2015
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11 **Abstract** We investigated the horizontal and vertical distribu-
12 tion of heavy metals (Hg, Pb, Zn, Cu, Cd, As, Ni, and Cr) in
13 soils in the water source protection zone for Shanghai to study
14 the origins of these metals, their connections with urbaniza-
15 tion, and their potential risk posed on the ecosystem.
Q2 16 Determination of metal concentrations in 50 topsoil samples
17 and nine soil profiles indicated that Hg, Pb, Zn, and Cu were
18 present in significantly higher concentrations in topsoil than in
19 deep soil layers. The spatial distributions of Hg, Pb, Zn, and
20 Cu and contamination hotspots for these metals in the study
21 area were similar to those near heavy industries and urban
22 built-up areas. Emissions from automobiles resulted in in-
23 creased soil concentrations of Cu, Pb, and Zn along roadsides,
24 while high concentrations of Hg in the soil resulted from re-
25 cent atmospheric deposition. Calculation of the potential

ecological risk indicated that the integrative risk of these 26
heavy metals in most areas was low, but a few sites surround- 27
ing high density of factories showed moderate risks. 28

Keywords Risk assessment · Land use · Vertical distribution · 29
Soil profile · Suburban area · Groundwater · Heavy metals · 30
Potential ecological risk 31

Introduction 32

To fulfill the resource needs of rapid industrialization and 33
urbanization, heavy metals have been mined from the 34
Earth's crust directly or with associated minerals. Since these 35
metals are minor components of fossil fuels and industrial 36
products, they tend to accumulate in the biosphere, resulting 37
in increased heavy metal content in soil, water, and the atmo- 38
sphere (Ajmone-Marsan and Biasioli 2010). Heavy metals in 39
the environment are non-biodegradable and subject to bioac- 40
cumulation (Wong et al. 2006). At natural background levels, 41
most heavy metals are beneficial to natural biota but become a 42
threat when they occur at high concentrations (Okuda et al. 43
2008; Wong et al. 2006). In China, many cities and counties 44
are facing the challenge of heavy metal pollution, particularly 45
of urban and agricultural soils (Cheng et al. 2014; Wei and 46
Yang 2010). 47

Urban areas are the geographical center of metal 48
emissions. Road networks and industrial plants are the 49
two main sources of heavy metal emissions in urbanized 50
and peri-urban areas (Chen et al. 2010; Dayani and 51
Mohammadi 2010; Wong et al. 2006). Previous studies 52
have shown that population density, traffic volume, in- 53
dustrialized land use, and timing of urbanization are 54
positively correlated with metal concentrations in sur- 55
rounding soils (Argyaki and Kelepertzis 2014; Guney 56

Responsible editor: Philippe Garrigues

Q3 **Summary** We investigated the spatial and vertical distribution of heavy
metals in soil in the Shanghai water supply area to determine the sources
of contamination and the impacts of urbanization.

Electronic supplementary material The online version of this article
(doi:10.1007/s11356-015-5745-3) contains supplementary material,
which is available to authorized users.

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57 et al. 2010; Peng et al. 2013). Soils in suburban and
 58 rural areas surrounding cities can be affected by atmo-
 59 spheric deposition of heavy metals originating from the
 60 city (Wei and Yang 2010; Wu et al. 2010). Due to high
 61 persistence of metals in soils, analyses can provide in-
 62 formation on historical metal contamination, e.g., asso-
 63 ciated with ancient human activities and former indus-
 64 trial activities (Chen et al. 2005; Chen et al. 2010; Xia
 65 et al. 2011; Zhang et al. 2005). Knowledge on the dis-
 66 tribution of heavy metals in soil is therefore essential
 67 for addressing the connections between metal accumula-
 68 tion and anthropogenic influences.

69 Heavy metals in soils are the result of both natural
 70 and anthropogenic processes. To determine the relative
 71 contributions of the two sources, previous studies have
 72 estimated natural background values, measured in re-
 73 mote pristine soils or deep soil layers (Biasioli et al.
 74 2006). Widespread atmospheric deposition of heavy
 75 metals originates from anthropogenic sources, but phys-
 76 ical disturbance of upper soil layers can affect the reli-
 77 ability of background value measurements (Wong et al.
 78 2006). It should be noted that background concentra-
 79 tions of heavy metals cannot be presented as single
 80 values but as a range of values that vary according to
 81 the geological and mineralogical characteristics of the
 82 parent material (Reimann and Garrett 2005). For these
 83 reasons, the single ratio of metal concentration to aver-
 84 age background value is not suitable for determining the
 85 influence of anthropogenic sources on heavy metal ac-
 86 cumulation (Reimann and Caritat 2000). It is therefore
 87 necessary to undertake statistical analysis and geograph-
 88 ical mapping in order to study the impacts of anthropo-
 89 genic activities on heavy metal accumulation in soils.

90 The metropolitan Shanghai, which has a total popu-
 91 lation of over 23 million, is facing rapid industrializa-
 92 tion and urbanization. Previous works had reported con-
 93 siderably increases of heavy metal contents in the urban
 94 soil and industrialized suburban soil of the city (Chen
 95 et al. 2012; Shi et al. 2008). The protection zones for
 96 centralized drinking water sources are essential for the
 97 health of residents and are located away from the urban
 98 areas. However, the rapid urban sprawl of Shanghai has
 99 extended the sphere of urban influence, which may be
 100 threatening the safety of drinking water sources. For this
 101 reason, we investigated the horizontal and vertical dis-
 102 tribution of heavy metals in soils along the upper
 103 reaches of the Huangpu River, the water source for
 104 metropolitan Shanghai. Our aims were (1) to determine
 105 the factors that cause and affect heavy metal accumula-
 106 tion in this area. (2) to examine the impacts of urbani-
 107 zation and industrialization on heavy metal accumula-
 108 tion in the area, and (3) to identify the potential eco-
 109 logical risks of heavy metal accumulation in the soils.

Method and analysis

Sampling

Dianshan Lake is regarded as the source of the upper Huangpu
 River, which runs through rural and suburban Shanghai and
 finally into the downtown area of the city (total length 75 km;
 Fig. 1). Since this river represents the major water source for
 Shanghai, many water supply points are located along its up-
 per reaches, such as the Changqiao waterworks, the Minhang
 waterworks, the second Qingpu waterworks, and the second
 Songjiang waterworks. To protect drinking water quality, the
 government has delineated a water source protection zone
 along the upper reaches of Huangpu River in which heavily
 polluting enterprises are prohibited. However, hundreds of
 factories already exist in this area and urban Shanghai has
 since expanded into the protection zone (Fig. 2).

To investigate the spatial distribution of heavy metals in the
 area, we divided the drinking water source protection zone
 into 5 km×5 km grids and randomly selected sites in each
 grid to obtain a total of 50 soil samples (Fig. 1). Each sample
 comprised five pooled subsamples of topsoil (0–20 cm) taken
 within 100 m² of the sampling site. The land use at the sam-
 pling sites was classified during the sampling process accord-
 ing to potential metal emission sources present as rural (*n*=33),
 roadsides (*n*=10), villages (*n*=4), and forest (*n*=3). The rural
 sites were further subdivided into two categories, namely farm-
 land sites (*n*=14) and countryside sites (*n*=19), with the latter
 having at least one industrial factory located within 500 m.

Soil profile samples were taken from six randomly
 selected rural sites (three farmland sites and three coun-
 tryside sites) to represent the vertical distribution of
 heavy metals in soils influenced/not influenced by con-
 tamination from industrial activities. The groundwater in
 the area is at around 1 m depth. Each soil profile was
 divided into five layers, 0–20, 20–40, 40–60, 60–80, and
 80–100 cm, and five samples were taken from each,
 resulting in a total of 30 soil profile samples. In addition,
 three soil profiles (each from the 0–10-, 10–20-, and 20–
 40-cm layer) were taken from the forest sites to represent
 soil only under the influence of atmospheric deposition.

Chemical analysis

The soil samples were air-dried and then milled to pass
 through a sieve with 0.1 mm mesh. The concentrations of
 Cd, Pb, Zn, Cu, As, Ni, and Cr in the samples were determined
 according to USEPA standard methods (6010C-2007): The
 milled soils were digested using a four-acid mixture of HCl,
 HNO₃, HF, and HClO₄, and the metal concentrations in the
 digested extracts were determined by ICP-AES (Optima 2100
 DV, PerkinElmer, USA). For analyses of Hg, the soils were
 digested in an acid mixture of H₂SO₄, HNO₃, and K₂Cr₂O₇,

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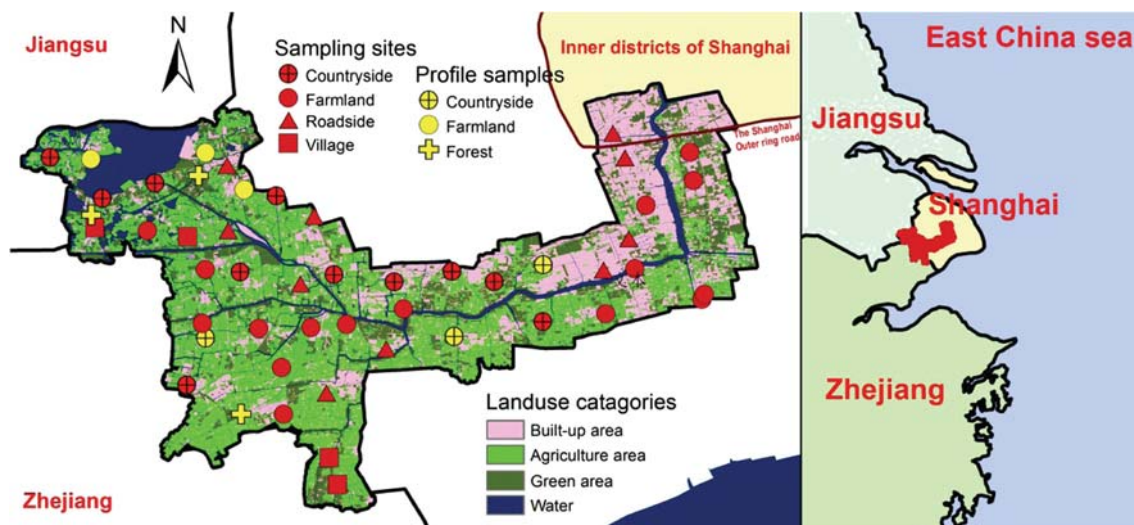


Fig. 1 The sampling sites and land use along the upper reaches of Huangpu River, the drinking water source protection zone of Shanghai

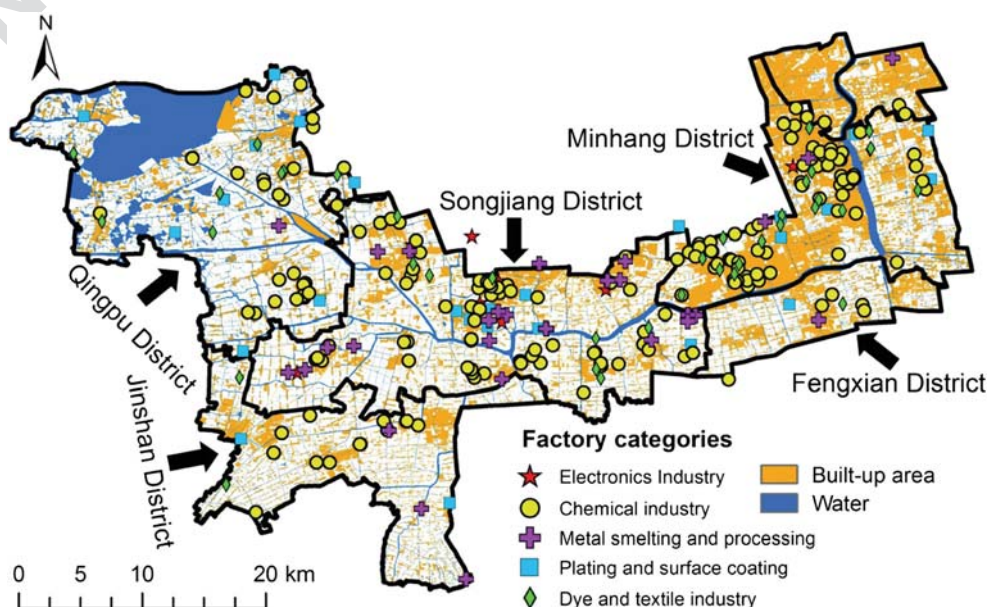
159 and the extracts were analyzed using a Mercury Automated
 160 Analyzer (MA-3000, Nippon Instruments Corporation,
 161 Japan). National standard reference soil (Center for Certified
 162 Reference Materials, China), duplicate samples, and blank
 163 samples were analyzed simultaneously in each array of assays.
 164 The recovery rate was 85–115 % for the heavy metals in the
 165 standard reference materials. Soil pH was measured in a
 166 soil/deionized water suspension of 1:2.5 (w/v). Soil water content
 167 was determined by the gravimetric method and soil bulk density
 168 by drying a soil core of known volume. Soil organic matter (SOM)
 169 was determined using the potassium dichromate oxidation procedure.
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Statistical analysis

Statistical analyses of the data were conducted in SPSS
 (version 18.0), including hierarchical cluster analysis,
 analysis of variance (ANOVA) analysis, the Kolmogorov-Smirnov
 test (K-S test), and regression analysis. Before cluster analysis
 and ANOVA, the metal concentration data were log-transformed
 to obtain an approximation of normal distribution. The calculation
 of metal leaching potential was performed in Excel 2010, while
 ArcGIS 10.1 software was used for geographical mapping and
 Getis-ord G_i^* analysis.

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Fig. 2 Spatial distributions of factories and constructed area along upper reaches of the Huangpu River



182 **Calculation of potential ecological risk**

183 The Hakanson potential ecological risk index (RI) was used
 184 widely for a comprehensive assessment of the harmful effects
 185 of heavy metals in soil (Qu et al. 2014; Yuan et al. 2014; Zhou
 186 et al. 2014). The risk index was calculated using the following
 187 equation (Hakanson 1980):

$$RI = \sum_i^n E_r^i = \sum_i^n T_r^i \times \frac{C_t^i}{C_f^i} \quad (1)$$

190
 191 where C_t^i denotes the concentrations of the i th metal species in
 192 soil and C_f^i the reference concentration of the i th metal species
 193 in pristine soil, in the current study, taken as the mean metal
 194 concentrations in the deep soil layers (80–100 cm). T_r^i is the
 195 toxic response factor for the i th metal species, which was rec-
 196 ommended as 40 for Hg, 30 for Cd, 10 for As, 5 for Pb, Cu,
 197 and Ni, 2 for Cr, and 1 for Zn (Qu et al. 2014). E_r^i represents the
 198 individual potential ecological risk index for the i th metal spe-
 199 cies. RI is the potential ecological risk index for multiple heavy
 200 metals in soil. The E_r^i can be grouped into four levels, and the
 201 RI is classified into five grades of risks (Table S2).

202 **Results and discussions**

203 **Metal concentrations and comparisons**

204 Heavy metal concentrations in topsoil along the upper reaches
 205 of the Huangpu River, a major water supply of metropolitan
 206 Shanghai, are shown in Table 1. The mean concentration of
 207 Hg, Pb, Zn, Cu, Cd, As, Ni, and Cr in the soils was 0.14, 21.5,
 208 87.7, 32.3, 0.19, 6.1, 30.9, and 66.1 mg/kg, respectively. The
 209 statistical distributions of Hg, Pb, Zn, and Cu showed higher
 210 values of skewness and kurtosis than those of Cd, As, Ni, and

Cr (Table 1). This suggests that the natural concentrations of 211
 Hg, Pb, Zn, and Cu in a few samples were significantly ele- 212
 vated by human activities. The metal concentrations observed 213
 were comparable to those reported for suburban soils in 214
 Beijing (HU et al. 2006) but lower than those observed in 215
 the industry-based suburban area of Wuxi (Zhao et al. 2007). 216
 Compared with two water source protection zones, in Beijing 217
 and the Pearl River delta, the Shanghai water source protec- 218
 tion zone had significantly higher concentrations in soil of all 219
 heavy metals except As (Table 2). These differences in metal 220
 concentrations between regions may be partly caused by the 221
 types of parent soil material, especially changes in median 222
 values. However, anthropogenic causes cannot be ignored, 223
 since the urban sprawl of Shanghai had reached the water 224
 source protection area for the city. 225

Table 1 shows the average background levels of heavy 226
 metals in Shanghai soils according to the National 227
 Environmental Monitoring Center of China. The mean and 228
 median concentrations of Hg, Zn, Cu, Cd, Ni, and Cr in deep 229
 soil layers of our study area were comparable to the average 230
 background values for Shanghai, but those of Pb and As were 231
 lower (Table 2). Differences in the mean and median concen- 232
 trations of Cd, As, Ni, and Cr between topsoil and deep soil 233
 layers in the study area were minor, suggesting that these 234
 metals originated from parent rock and were less affected by 235
 modern industrial activities. However, the large differences in 236
 mean concentrations of Hg, Pb, Zn, and Cu between surface 237
 soils and deep soils indicate anthropogenic emissions of those 238
 metals in the upper reaches of the Huangpu River (Table 2). 239
 Hierarchical cluster analysis to improve understanding of 240
 heavy metal accumulation patterns revealed that the eight 241
 metals studied here were grouped into two major clusters in 242
 topsoil, namely [Pb + Zn + Cu + Hg] and [Cr + Ni + As + Cd] 243
 (Fig. 3). This grouping suggests the heavy metals in this area 244
 have two major accumulation patterns. In the first group, [Pb 245

t1.1 **Table 1** Statistical parameters for metal concentrations in topsoil along the upper reaches of the Huangpu River

t1.2		Hg (mg/kg)	Pb (mg/kg)	Zn (mg/kg)	Cu (mg/kg)	Cd (mg/kg)	As (mg/kg)	Ni (mg/kg)	Cr (mg/kg)	θ_w (I_{water}/I_{soil})	pH	ρ (kg/l)	SOM (%)	
t1.3	Mean	0.14	21.5	87.7	32.3	0.186	6.1	30.9	66.1	24.48	7.47	1.27	3.01	
t1.4	Std. deviation	0.09	8.65	23.31	9.24	0.068	1.9	4.37	10.07	9.25	0.85	0.16	1.06	
t1.5	Minimum	0.04	9.36	52.6	13.9	0.050	3.6	22.5	47.7	6.53	4.15	0.99	0.36	
t1.6	Percentiles	25 %	0.07	16.38	75.08	27	0.150	4.8	28.15	58.33	16.73	7.4	1.17	2.4
t1.7		50 %	0.12	18.75	83.15	30.8	0.180	5.6	30.65	66.6	23.15	7.69	1.26	2.88
t1.8		75 %	0.16	24.8	95.18	37.78	0.223	6.8	33.95	71.73	32.26	7.92	1.37	3.57
t1.9		90 %	0.26	33.18	115.7	45.26	0.279	9.9	36.76	82.59	37.94	8.05	1.56	4.55
t1.10	Maximum	0.48	53	177	70.4	0.340	10.8	41.8	86.7	46.43	8.29	1.61	6.02	
t1.11	Skewness	1.79	1.73	1.69	1.42	0.25	1.04	0.14	0.08	0.26	-2.57	0.48	0.47	
t1.12	Kurtosis	3.58	3.55	4.11	4.92	0.28	0.31	-0.12	-0.52	-0.68	6.72	-0.34	1.89	
t1.13	Backgrounds (CNEMC 1990)	0.043	21	74.7	24.4	0.133	9	29.9	63.4					

Table 2 Mean (median) concentrations of heavy metals reported in soils along different reaches of the Huangpu River, Shanghai, in the Pearl River Delta, Guangdong, and the Miyun Reservoir, Beijing

t2.1	t2.2	t2.3	t2.4	t2.5	t2.6	t2.7	t2.8	t2.9	t2.10			
Location	Land use	N	Depth	Hg (mg/kg)	Pb (mg/kg)	Zn (mg/kg)	Cu (mg/kg)	Cd (mg/kg)	As (mg/kg)	Ni (mg/kg)	Cr (mg/kg)	References
Upper reaches of Huangpu River	Water source protection zone	50	0–20 cm	0.14 (0.12) ^a	21.5 (18.8)	87.7 (83.2)	32.3 (30.8)	0.186 (0.180)	6.1 (5.6)	30.9 (30.7)	66.1 (66.6)	Present study
Middle reaches of Huangpu River	Urban area	167	0–20 cm	0.066 (0.058)	15.7 (16.0)	69.3 (65.7)	23.2 (22.0)	0.155 (0.165)	6.87 (6.96)	29.5 (29.9)	59.7 (58.2)	Cheng et al. (2014)
Middle reaches of Huangpu River	Urban area	42	150–180 cm	0.358 (0.406)	103.6 (73.5)	244 (128)	63.8 (47.9)	1.091 (0.37)	9.5 (8.7)	36.2 (17.9)	114 (96)	Shi et al. (2008)
Lower reaches of Huangpu River	Industrialized suburban area	273	0–10 cm	–	70.7	301.4	59.3	0.52	–	31.1	107.9	Chen et al. (2012)
Pearl River delta, Guangdong, south China	Water source protection zone	48	0–20 cm	0.15	40.24	149.8	39.58	0.19	7.41	–	111.4	Hu et al. (2013)
Miyun Reservoir, Beijing, north China	Water source protection zone	227	0–10 cm	0.06	34.2	47	10.6	0.12	10.4	7.4	39.1	Luo et al. (2010)
		95	0–20 cm	–	11 (10)	64 (55)	23 (18)	0.05 (0.05)	6.5 (6.7)	18 (16)	40 (33)	

^a Numbers in brackets indicate median values

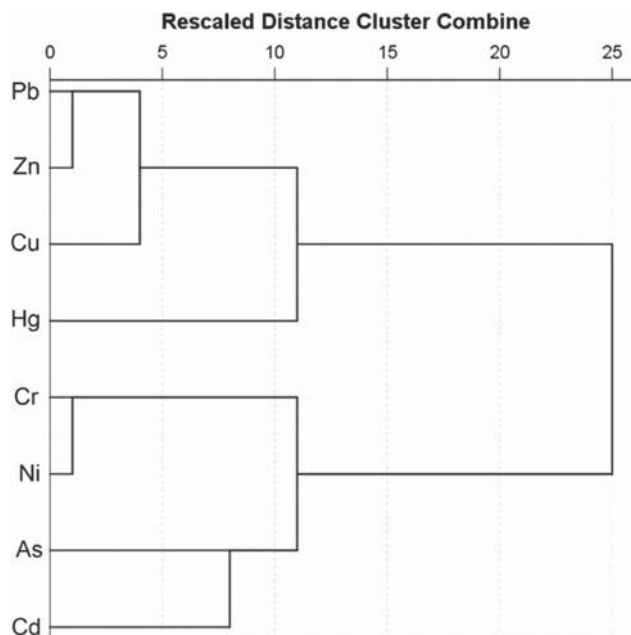


Fig. 3 Dendrogram of heavy metals in soils along the upper reaches of Huangpu River using hierarchical cluster analysis

+ Zn + Cu + Hg], concentrations in topsoil were higher than in deep soil layers, indicating exterior inputs from anthropogenic sources. On the other hand, the topsoil concentrations of the second group, [Cr + Ni + As + Cd], showed normal distributions and the differences between topsoil and deep soil were minor, suggesting that the concentrations of these metals were not influenced by human activities. The sources, distribution, and leaching potential of Hg, Pb, Zn, and Cu in this area should be further investigated.

Impacts of urbanization on metal distribution

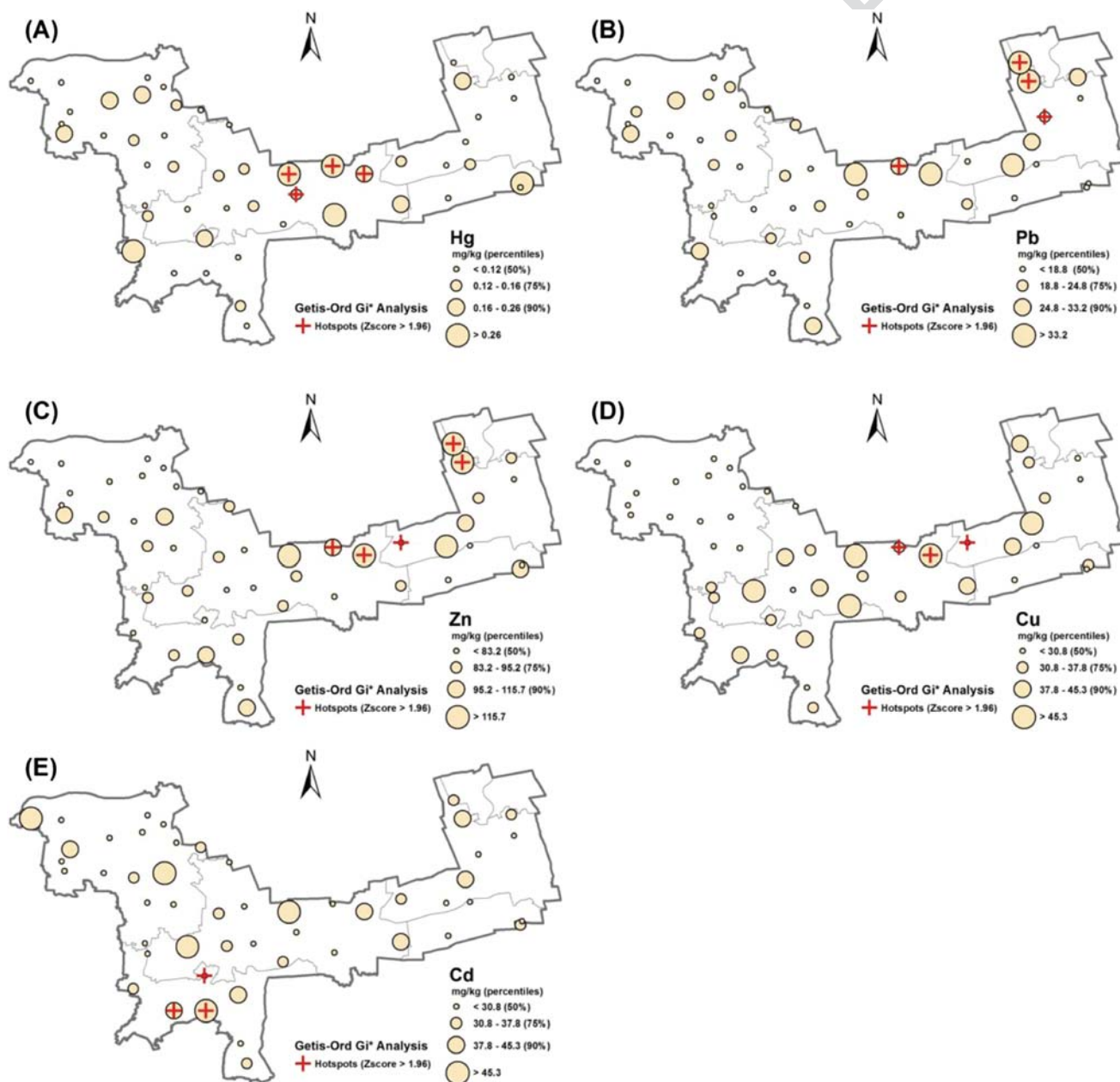
The Huangpu River flows across rural, suburban, and urban areas of Shanghai. Previous studies have reported on contamination levels of heavy metals in the soils around the middle and lower reaches of this river (Table 2). Compared to the results of the present study, these studies have indicated that the soil concentrations of Hg, Pb, Zn, Cu, Cd, and Cr had increased to double and six times in the urban areas of Shanghai (Cheng et al. 2014; Shi et al. 2008). Minor changes in Ni concentrations were noted in soils along the upper-middle-lower reaches of the Huangpu River. Chen et al. (2012) reported concentrations of Pb, Zn, Cu, and Cr in the soils of Baoshan District, a heavily industrialized suburban area of Shanghai near the lower reaches of the Huangpu River that were significantly higher than those found in the present study. Overall, soil metal concentrations vary dramatically along the upper-middle-lower reaches of the Huangpu River, suggesting that urbanization and industrialization are the main causes of metal contamination in the city. We found that the soils along the upper reaches of the Huangpu River

275 had the lowest mean concentrations of heavy metals of all
 276 regions in the city. However, the impacts of urbanization and
 277 industrial activities on heavy metal distribution inside this
 278 region provide cause for concern, because relatively high con-
 279 centrations in even a few samples can threaten water quality.

280 Figure 4 shows the spatial distribution of Hg, Pb, Zn, Cu,
 281 and Cd measured in soils along the upper reaches of Huangpu
 282 River. The size of the graphic symbols used in the diagram
 283 is proportional to metal concentration within different percen-
 284 tiles. Getis-ord G_i^* analysis was used to mark the hotspots
 285 of metal contamination, by means of studying the spatial au-
 286 tocorrelation of sampling sites (Fig. 4). Each hotspot indicates
 287 a sampling site surrounded by several highly polluted

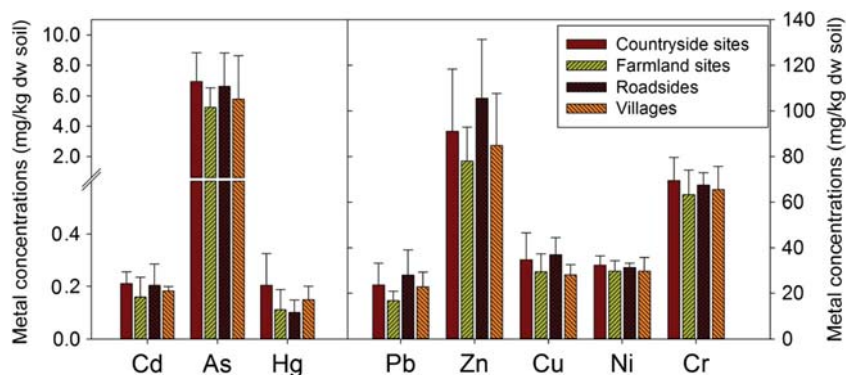
288 sampling sites that were aggregated to form an area of high
 289 metal contamination. Spatial differences in parent soil mate-
 290 rials may have caused natural fluctuations in metal concentra-
 291 tions in the soils, but the highlighted hotspots and the sam-
 292 pling sites with high metal concentrations (i.e., with the largest
 293 symbols in Fig. 4) are probably associated with human activi-
 294 ties. Hotspots and sites with high concentrations of Hg, Pb,
 295 Zn, and Cu tended to aggregate in west Minhang district and
 296 east Songjiang district (Fig. 2), while the high Cd concentra-
 297 tion areas were randomly distributed along the upper reaches
 298 of the Huangpu River (Fig. 4e).

299 Industrial activities and urban transportation were impor-
 300 tant emission sources of heavy metals. Figure 2 shows the



Q4 **Fig. 4** Spatial distributions of Hg, Pb, Zn, Cu, and Cd in the soils along upper reaches of the Huangpu River

Fig. 5 Concentrations of heavy metals in soils under different land utilizations



301 spatial distributions of heavy industries and urban built-up
 302 areas along the upper reaches of the Huangpu River. The
 303 density of built-up area can be used as an indicator of the
 304 extent of urbanization and transportation. The factories in
 305 the study area were preliminarily classified into five categories:
 306 electronics, chemicals, metal smelting and processing,
 307 plating and surface coating, and dye and textiles. The spatial
 308 distribution of Hg, Pb, Zn, and Cu in the soils appeared to
 309 show some similarities to the positions of industries and
 310 built-up areas (Figs. 2, 3, and 4). A high density of factories
 311 and high levels of urbanization were mainly found in west
 312 Minhang and east Songjiang districts, in which most hotspots
 313 and high contamination sites of Hg, Pb, Zn, and Cu also
 314 occurred.

315 **Impacts of land utilization**

316 We classified the sampling sites into four land use types ac-
 317 cording to the distribution of metal emission sources, namely
 318 farmland sites, countryside sites, roadsides, and villages.
 319 Differences in mean concentrations of the eight metals studied
 320 were noted in these four land use areas (Fig. 5). The highest
 321 mean concentrations of Cu, Pb, and Zn were found in roadside
 322 soils (37.7 mg/kg for Cd, 27.9 mg/kg for Pb, and 105.5 mg/kg
 323 for Zn), while the corresponding mean concentrations of Cu,
 324 Pb, and Zn in farmland soils were significantly lower
 325 (ANOVA: $p < 0.05$), 29.3, 16.7, and 78.0 mg/kg, respectively.
 326 This suggests that automobile traffic is a significant source of
 327 these metals. Wear on brakes, tires, and other vehicle

328 components can release large quantities of Cu and Zn into
 329 the environment (Johansson et al. 2009), while the Pb in road-
 330 side soils may be a legacy of burning leaded fuel (Biasioli
 331 et al. 2006).

332 The mean concentrations of Hg, Pb, Cd, and As were
 333 0.201, 23.6, 0.211, and 6.96 mg/kg, respectively, in the coun-
 334 tryside soils and 0.111, 16.8, 0.159, and 5.2 mg/kg, respec-
 335 tively, in the farmland soils (Fig. 5). ANOVA showed that the
 336 differences were significant ($p < 0.05$). We distinguished two
 337 categories (farmland and countryside sites) in the rural soils
 338 sampled, which were and were not, respectively, under the
 339 direct influence of industry activities. Wastewater and gas dis-
 340 charges from industries may enter surrounding soils through
 341 atmospheric deposition and/or irrigation, thus increasing soil
 342 metal concentrations. For example, chemical and metallurgy
 343 industries are responsible for high emissions of Hg, Pb, Cd,
 344 and As into the surrounding environment by these routes
 345 (Ajmone-Marsan and Biasioli 2010; Cazier et al. 2011). The
 346 Cd and Pb found in the Huangpu River soils studied here may
 347 have originated from plating and coating industries. The elec-
 348 tronics industry is also responsible for metal emissions
 349 (Ajmone-Marsan and Biasioli 2010).

350 **Vertical distribution of heavy metals in the soils**

351 The concentrations of heavy metals in different soil layers are
 352 commonly used to investigate their accumulation history and
 353 leaching processes. Figure 6 shows the vertical distribution of
 354 heavy metals in the soil profiles from the three countryside

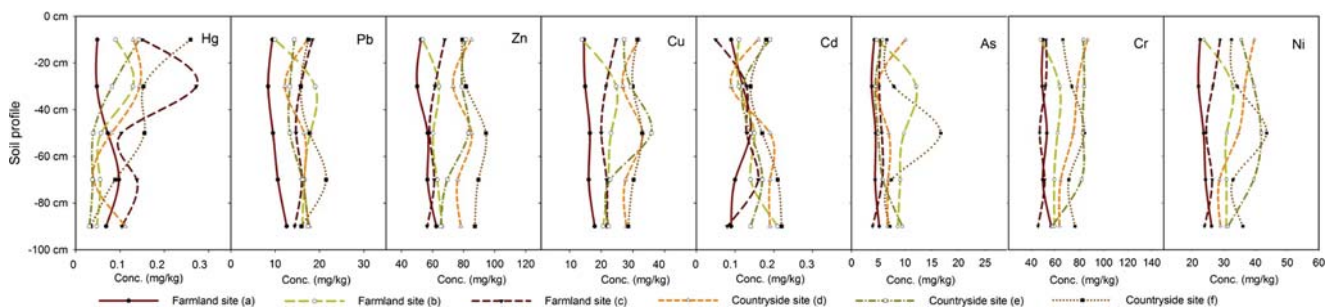


Fig. 6 Profiles of heavy metals in rural soils along the upper reaches of Huangpu River

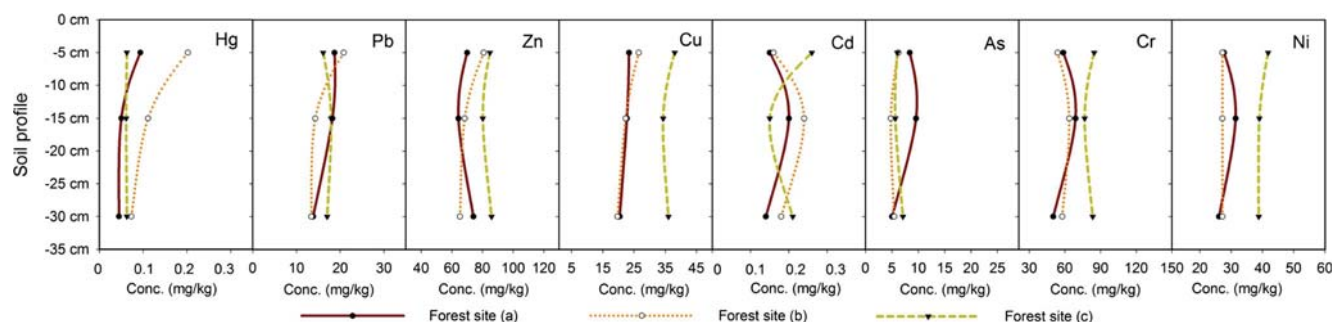


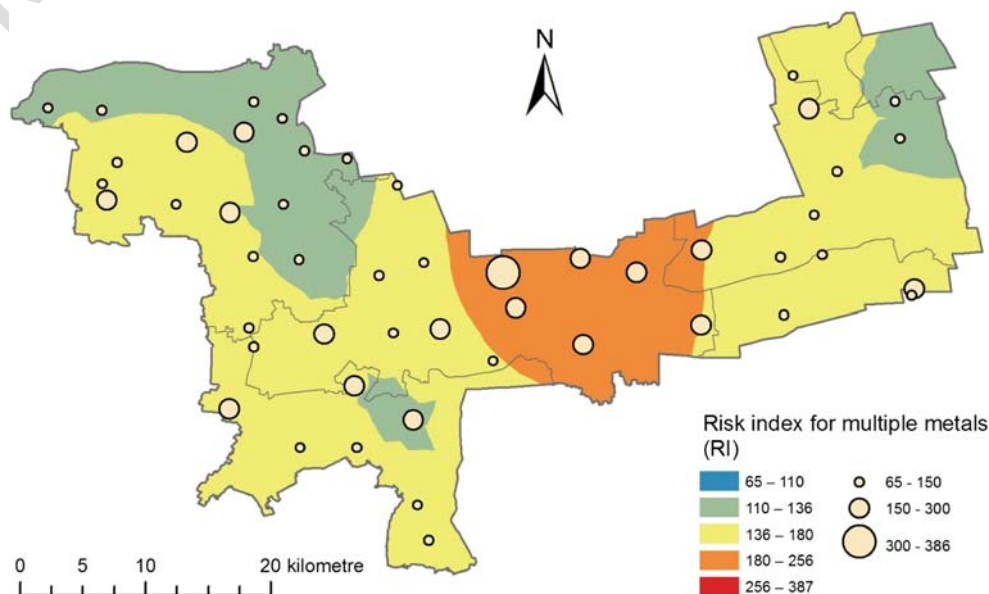
Fig. 7 Profiles of heavy metals in forest soils along the upper reaches of Huangpu River

355 sites near industries and the three control farmland sites that
 356 were beyond the direct influence of industrial activities. The
 357 metal concentrations in most soil layers were slightly higher at
 358 the three countryside sites than at the farmland sites but fluctu-
 359 ated with depth in the soil. There was no clear difference in
 360 metal concentrations between the top and bottom soil layers
 361 except for Hg, the concentrations of which showed a decreas-
 362 ing gradient with depth. This indicates accumulation of Hg in
 363 surface soil as a result of recent urbanization and industrializa-
 364 tion. At one farmland site (site (c) in Fig. 6), the concentra-
 365 tions of Hg, Pb, Zn, Cu, As, Cr, and Ni in the 20–40-cm soil
 366 layer were considerably higher than those in the 0–20-cm soil
 367 layer, suggesting that the surface soil at this site might have
 368 been recently transported there from elsewhere for agricultural
 369 purposes. At one countryside site (site (f) in Fig. 6), the con-
 370 centrations of Zn, Cu, As, and Ni showed abnormal increases
 371 at 40–60 cm depth, indicating a buried soil layer with histor-
 372 ical contamination. Therefore, the major factors affecting met-
 373 al concentrations in soil layers were fluctuations in back-
 374 ground values and past physical disturbance. The impacts of
 375 soil solute leaching on metal concentrations in deep soil layers

were probably minor, due to the strong adsorption of heavy
 metals to soil particles (Chen et al. 2009).

Forest soil profiles were sampled at remote sites (Fig. 7),
 far from industries and urban traffic, where atmospheric de-
 position was the most likely path by which heavy metals
 would have entered the soil (as opposed to farmland, which
 was under direct human interventions such as irrigation). The
 mean concentration of Hg, Pb, Zn, Cu, Cd, As, Ni, and Cr in
 the 20–40-cm layer of forest soils was 0.06, 14.7, 75.1, 25.4,
 0.18, 5.86, 30.7, and 63.7 mg/kg, respectively. These were
 similar to the concentrations in the 80–100-cm layer at
 farmland sites but lower than those in the deep soil layer
 in urban Shanghai (Table 2), indicating the natural back-
 ground of heavy metals in this region. No significant
 differences in Pb, Zn, Cu, Cd, As, Cr, and Ni concentra-
 tion were found between the topsoil and deeper soil
 layers at the forest sites (ANOVA: $p > 0.05$). A high Hg
 content was observed in the first (upper) soil layer of
 forest land and decreased rapidly with increasing soil
 depth, which suggests that atmospheric deposition was
 the major input of Hg in this region.

Fig. 8 Spatial distribution of the potential ecological risk in soils along the upper reaches of Huangpu River



397 Potential ecological risk assessments

398 The potential ecological risk index for individual metal spe-
 399 cies (E_r^i) and multiple metals (RI) are summarized in Table S1,
 400 the mean of which follows the gradient of $RI > Hg > Cd > As$
 401 $> Cu > Pb > Ni > Cr > Zn$. According to the classification of
 402 E_r^i , only Hg and Cd had their concentrations exceeding the
 403 moderate risk level (Table S2). Hg in 6 and 30 % of the
 404 sampling sites had high and considerable risks, respectively,
 405 while the E_r^i of Pb, Zn, Cu, As, Ni, and Cr in all the samples
 406 showed low risk levels. These results suggested higher eco-
 407 logical risks for Hg and Cd than the other metals, especially
 408 Hg should be of concern in terms of the high ecological hazard
 409 it may pose to the surface soil in the water source protection
 410 area.

411 The RI value is the integrative risk index of multiple con-
 412 taminants, ranged from 65.5 to 386.5 with a mean value of
 413 151.6 in the study area (Table S1). The threshold values of RI
 414 between the low, moderate, and considerable risk levels are
 415 150 and 300, respectively, based on which the sampling sites
 416 are classified into three groups, namely low risk (64 %), mod-
 417 erate risk (34 %), and considerable risk (2 %) (Fig. 8). The east
 418 Songjiang district is marked as a hotspot area of RI, in which
 419 high density of factories can be found (Fig. 8). Meanwhile, the
 420 west Minhang district with the highest level of urbanization
 421 and Pb/Zn contents however shows relatively lower RI. The
 422 spatial distribution of RI is similar to the distribution of Hg
 423 (Fig. 4), because the E_r^i of Hg contribute on average 55 % of
 424 the risk index to RI in the study area (Fig. S1). The differences
 425 of E_r^i between Hg and the other metals mainly resulted from
 426 their toxic response factors (T_r^i) and abnormally high concen-
 427 tration at a few sampling sites (Zhou et al. 2014). Therefore,
 428 RI reflects the risk of high toxic metals that are released by
 429 industrial activities in this region.

430 Conclusions

431 This analysis of horizontal and vertical distributions of heavy
 432 metals in soils along the upper reaches of the Huangpu River
 433 (the drinking water source for Shanghai) revealed significant-
 434 ly elevated concentrations of Hg, Pb, Zn, and Cu in topsoil,
 435 while the concentrations of As, Ni, and Cr were close to nat-
 436 ural background levels. Spatial distributions and Hg, Pb, Zn,
 437 and Cu hotspots in the soils coincided with the location of
 438 heavy industries and built-up areas. Emissions from factories
 439 have resulted in increased levels of Hg, Pb, Cd, and As in
 440 surrounding countryside soils. Elevated concentrations of
 441 Cu, Pb, and Zn were also found in roadside soils.
 442 Examination of the metal distributions in soil profiles sug-
 443 gested that the elevated Hg concentrations were probably the
 444 result of recent urbanization and industrialization in the sur-
 445 rounding areas. On calculating the potential ecological risks,

we concluded that the presence of heavy metals in soils in the 446
 study area poses moderate risk to the east Songjiang district, 447
 which is associated with the industrial activities in this region. 448

Acknowledgments We gratefully acknowledge financial support pro- 449
 vided by the National Natural Science Foundation of China (Grant No. 450
 41401588). 451

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